

The Effect of Seasonal Rainfall on Groundwater Quality in North Kuwait

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ABSTRACT: Water is substantial natural resources on earth which cater for all human endeavors. Groundwater is the most essential and reliable source of fresh water in arid zones like north Kuwait. This study evaluates the effect of rainfall precipitation on groundwater characteristics in north Kuwait. The representative samples of groundwater were collected from 50 water points, in addition to rain water and Arabian Gulf water samples. The chemical analyses of the collected water samples were performed in the Laboratories of the MEW. The results showed that the groundwater moves generally from the SW to the NE towards the Arabian Gulf. Direct recharge through the rainfall, faults and lateral flow from the adjacent aquifers are the acceptable source of the recharge. The salinity change suggests that the fresh water-saline water interface, as defined by the 1500 ppm contour line, has moved towards the centre of the fresh water lens. In some areas up-gradient of the fresh water recharge by rainfall, the extent of the fresh water lens has actually increased slightly in the down-gradient areas. In addition, the relation between TDS and different major ions are statistically analyzed. Four principal components were extracted from chemical data to explain the major sources and processes responsible for chemical characteristics of groundwater. Cluster analysis showed that silicate weathering, agricultural runoff (fertilizer input), municipal wastewater infiltration play a vital role in the groundwater pollution. It is concluded that the majority of the samples are good to permissible for drinking purpose. It is highly recommended that the environmental problems related to the damaged during the 1991 Gulf War are badly in need of treatment through continuous projects as fast as possible before any urbanization development or any integrated water management.

KEYWARDS: Hydrogeology, Geochemical processes, Al-Raudhatain and Umm Al-Aish, north Kuwait.

Date of Submission: 27-07-2018

Date of acceptance: 11-08-2018

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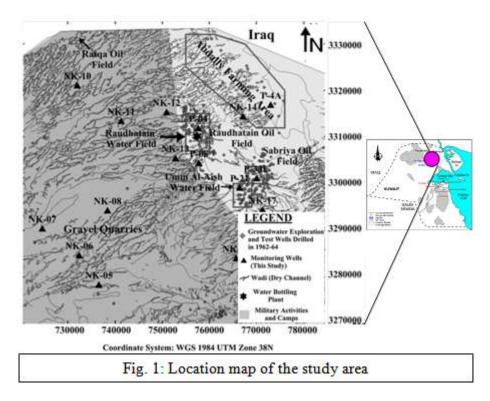
I. INTRODUCTION

Kuwait state is situated in an arid coastal region characterized by high temperatures, low humidity, sparse precipitation rates, and high evaporation and evapotranspiration rates with no rivers or lakes. The acute lack of natural water resources of the country has created a unique water sector in Kuwait, that is, about 90% of the potable water production is coming from the unconventional resource of seawater desalination. Although this unusual situation has been maintained for decades and probably will continue into the foreseeable future, the complete dependence on a highly industrialized source for freshwater has serious drawbacks, one of which is the lack of secured strategic water reserve. As such, the usable water accumulation in northern Kuwait stands out as a potential sizable emergency reserve (Al-Senafy et al, 2013). The infiltration of the rainwater/runoff from the occasional rainstorms of Kuwait is known to produce freshwater lenses in the natural depressions of northern Kuwait (Parsons, 1964). The infiltration of this accumulated water over thousands of years has given rise to the formation of freshwater lenses below these depressions and brackish water reserves surrounding them (Amitabha Mukhopadhyay et al. 2016). Nonetheless, freshwater accumulations are not confined to the depressions, but extend however in lesser volumes, to the formations underlying the wadis that carry the surface base flows. During the early exploration efforts, potential usable water accumulations, i.e., total dissolved solids (TDS) <5000 mg/l, of different volumes have been reported in many parts of northern Kuwait (Parsons, 1964). While the main two depressions (Raudhatain and Umm Al-Aish) have been evaluated extensively, most of the

other identified/potential usable water accumulations have been given next to no attention. This was mainly due to their irrelevance in the pursuit of freshwater bodies large enough for economical exploitation. However, in the context of securing a strategic emergency reserve for the country, these relatively smaller accumulations are significant assets. The lack of information on the location, extent and TDS content of these accumulations as well as the recharge rates to the watershed are deficiencies that prevent preparation of emergency utilization plans for these resources. In this context to address these deficiencies, this study was conducted with the overall aim of exploring the effect of seasonal rainfall on groundwater quality of the Shallow Groundwater Resources in Northern Kuwait (SGRNK).

1-1 Site description and climate.

The study area is located in the northern part of Kuwait state. It lies far from Kuwait city by about 55 Km in NW direction and west of Arabian Gulf by about 25 Km (Fig.1).



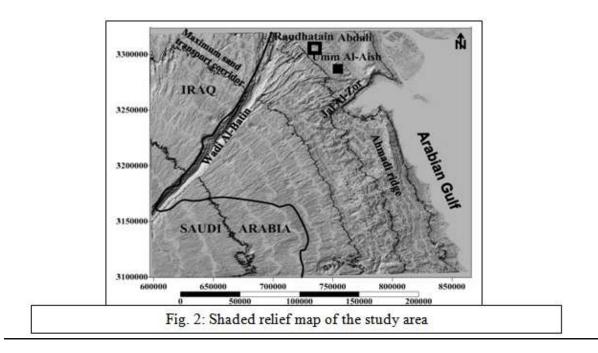
It is limited between latitudes 3270000 and 3330000 due North and longitudes 725000 and 780000 due East with an area of 1650 Km^2 . Al-Jahra table land area bounds SGRNK from south while from the east, Arabian Gulf forms the northern boundary Iraq-Kuwait border represents the northern boundary. The climate of north Kuwait can be divided into two main seasons, hot with temperature ranges between 46 °C and 50 °C during summer months and from 20 °C to Zero °C during winter months (November through March). The mean annual precipitation was about 101.02 mm, and the monthly average 9.6 mm while the mean daily Pan-A evaporation rate varied from 4.7 mm in January to 31 mm in July, with an average mean daily rate of 9.4 mm (Safar 1985). Moreover, the annual evaporation rate ranges between 8.8 mm and 9.8 mm while the relative humidity ranges from 31.1% to 38.7% with mean value of 34.66% (Table 1).

Serial No.	Year	Rainfall in (mm)	Evaporation in (mm)	Humidity in (%)
1	2012	126.7	9.4	31.1
2	2013	117.9	9.5	33.9
3	2014	73.7	9.5	38.7
4	2015	97.8	8.8	37.1
5	2016	89	9.8	32.5
Mean		101.02	9.4	34.66

Table 1: Climatic characteristics of the study area

1.2 Geomorphological settings

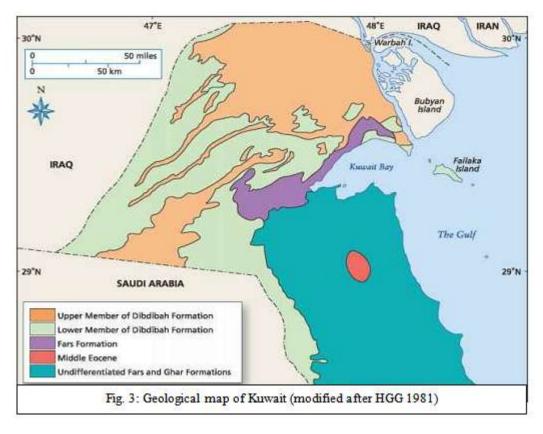
Geomorphological features play a vital part in accumulation of rainwater and irrigated drainage water in low depressions and consequently affect the soil water quality. To delineate the geomorphological setting of the study area, the TM images and other maps were rectified to the Universal Transverse Mercator (UTM) zone 38 projection and co-ordinate system, using a second order polynomial re-sampling scheme and bilinear interpolation (Lillesand & Kiefer, 1994). The shaded relief of the study area was generated from the DEM data and TM images (Fig. 2).



The resulted shaded relief represents elevations ranging from 0 to 300 m above mean sea level. The highest elevations occur in the SW with a gradual eastward decline towards the Gulf. The extracted geomorphological features in the study area include the NE-SW trending Wadi Al-Batin, an ancient riverbed (Al-Sarawi, 1980; Al-Sulaimi, Pitty, 1995 and Kwarteng et al. 2000), and the sand corridor. Otherwise, several geomorphological features with implications for groundwater exploration are readily observed. The most prominent include the Jal Al-Zor escarpment, the Ahmadi ridge, the Wadi Al-Batin valley, the Raudhatain Umm Al-Aish, and Abdali depressions, and the sand corridor.

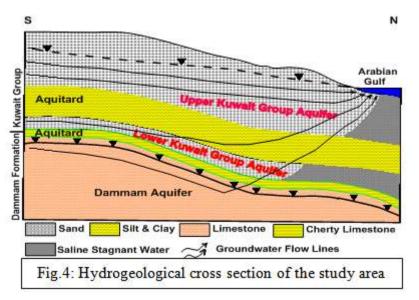
1.3 Geological settings

The geology of Kuwait consists of recent (Quaternary age) deposits and sediments unconformably overlying sedimentary rocks (Tertiary age) which host the groundwater aquifers. The regional dip of strata is about 2 m per km towards the north-east and this regular dip is interrupted by anticlines and other small structures which are present at the Raudhatain oil field. The Dibdibba Formation, confined to the northern part of Kuwait, comprises sands and gravels with minor clay and gypsiferous sandy clay beds. The details of the geology of rock units of Kuwait have been presented by Owen and Nasr (1958), Milton (1967), Fuchs et al. (1968), Burdon and Al-Sharhan (1968), Omar et al. (1981), Clarke (1988), Al-Sulaimi (1988), Amer et al. (1989), Al-Sulaimi and Pitty (1995), Mukhopadhyay et al. (1996), Al-Sulaimi and Mukhopadhyay (2000), Al-Sulaimi and Al-Ruwaih (2004) as well as the geologic maps and landsat images (El-Baz & Al-Sarawi 2000) (Fig.3). In addition, the paleo-drainage channels in SGRNK, which were formed in the Pleistocene (Al-Sulaimi et al., 1997), are carved in the Upper and Lower Dibdibba and the undifferentiated Fars and Ghar formations. Presently, they are filled with gravel and sand and are not readily observed on flat terrain where they are only manifested as micro-relief with the surroundings. Moreover, the relative abundance of paleo-drainage channels in the north and south-west is due to the underlying hard calcretic and gypcretic gravelly deposits of the Dibdibba Formation. Conversely, the paucity of wadis in the south is due to the friable sandstone of the Undifferentiated Fars and Ghar Formation, which was not as ideal for developing and preserving the drainage channels. The south-west-north-east trending drainage pattern closely follows the present relief variations (Al-Senafy et al. 2013).



1.4 Hydrogeological setting

Hydrogeologically, the two fresh water aquifers of interest are situated beneath the shallow, elongated depression of the Al-Raudhatain and Umm Al-Aish drainage system which consist of an extensive network of wadis and drainage lines (Yihdego and Al-Weshah 2016). The S-N hydrogeological cross-section shows that upper aquifer, known as the Kuwait Group aquifer, consists of silty, gravelly sand; and the lower one, known as the Dammam Formation, consists of chalky and dolomitic limestone (Fig. 4). The Kuwait Group aquifer is generally in an unconfined state, whereas the Dammam Formation is confined. Moreover, that the Kuwait Group aquifer is poor in hydraulic properties. The porosity ranges from 5% to 20% and hydraulic conductivity ranges from 17 to 71 m/day while the highest estimated transmissivity reaches 1998 m²/d and storativity reaches 0.00018 (GII, 2010). The groundwater table varies from approximately 90 m above mean sea level in the SW, and decreases to zero at the Gulf coast.



The flow of groundwater is from SW to NE. In general, the groundwater quality varies from brackish in the SW to highly saline in the NE region. Brackish groundwater with TDS of less than 3000 mg is extracted and used for irrigation and landscaping. The majority of the groundwater wells extracting brackish groundwater exist in the central and SW regions of Kuwait. Beyond the brackish groundwater fields, groundwater quality deteriorates rapidly with TDS exceeding 10,000 mg/1 and reaching levels as high as 100,000 mg/l in the northern and NE regions (Kwarteng et al. 2000).

On the other hand, in the north, the groundwater of SGRNK which was accidentally discovered in the early 1950s (Parsons Corporation, 1964), freshwater lenses of TDS

less than 1000 mg/1 are found floating on saline groundwater of TDS more than 100,000 mg/1. The boundary between the freshwater and saline water is diffuse, however, the density difference between the two results in stable configuration of the lenses. Extensive investigations were carried out in the early 1960s by Parsons Corporation (1964). The investigation concluded that fresh groundwater in SGRNK existed as lenses that were formed by infrequent infiltration of rainwater. The study estimated the total volume of freshwater in Raudhatain basin to be approximately 68,130 Ml. Senay (1977) estimated the safe yield from the Raudhatain field to be about 6.8 Ml/day. Fresh groundwater from the lenses has been extracted for potable purposes and to produce bottled mineral water since the 1960s. Freshwater extraction for potable purposes from these fields was discontinued after 1977 because of the development of large desalination plants. Presently, groundwater characteristics in SGRNK is very important for sustainable development.

II. MATERIALS AND METHODS

To study the seasonal change of groundwater quality after seasonal rainfall in SGRNK, 12 field trips during the period 2015-17 were carried out. During these field trips, 120 groundwater samples were collected in August, 2016 (pre-monsoon) and January, 2017 (post-monsoon) from dug-wells and bore-wells and locations were fixed by GPS (Oregon 600). Samples were taken in properly rinsed 250 ml polyethylene bottles (presoaked in acid wash for 24 h. and rinsed several times with distilled water). General parameters such as pH, EC, TDS and depth to water table were measured immediately at the time of sampling using a multi parameter ion meter (pH/Cond 340i SET 1). Moreover, the groundwater samples were analyzed at NAPESCO Environmental Laboratory in Kuwait, which is approved with KEPA for conducting all baseline sampling and analysis. NAPESCO laboratory has a fully fledged Quality Management System (QMS) with ISO 17025:2005 accreditation.

The physicochemical parameters were determined using the standard analytical methods (APHA 2005): hardness, alkalinity, Ca, CO₃, HCO₃ and Cl were analyzed with titrimetric method and Mg was determined with calculation method. Na and K was analyzed using flame photometer (Elico Model CL 378). F was estimated using ion analyzer (Thermo scientific Orion-4 star) with an ion-selective electrode. SO4, NO₃ and PO₄ were determined by spectrophotometry method (UV 3200 double beam spectrophotometer, Labindia). The analytical precision for the accurate measurements of ions was determined by calculating electrical neutrality (EN %) which is acceptable at ± 5 % (Appelo and Postma 1999). All the samples have EN % values within ± 5 % in pre- and post-monsoon.

Electrical Neutrality = Σ Cation + Σ Anion / Σ Cation - Σ Anion x100.....(1)

The analytical data obtained were processed for detailed geochemical and statistical analysis. Statistical analysis was carried out by Stat Soft STATISTICA 10 and Microsoft Excel-2010. Salinity mapping and hydrogeochemical facies distribution maps were prepared by Golden software Surfer 13. Total hardness and various water indices used to classify groundwater suitability for irrigation and industrial purposes are calculated using the formula:

Total Hardness = 2.5 Ca + 4.1 Mg (mg/l as CaCO3)....(2)

As a general, in the laboratory, the water samples were filtered through 0.45 μ m to separate suspended particles. Acidification (pH < 2) with concentrated nitric acid was performed on the filtered samples for heavy metals analysis using ICP (Inductively Coupled Plasma) at NAPESCO laboratory. The acid titration method was used to determine the concentration of bicarbonate HCO⁻³ (APHA, 1995). Trace elements (Fe and Mn) were analyzed using ion chromatography (Dionex DX-600). The comparison between salinity measurements at the end of the dry season (during September 2010) and in the middle of the rainy season (January 2011) is given in (Fig.5). In addition, the base case scenario of the groundwater quality in presented in the results of the routine analysis of 40 groundwater samples collected from SGRNK at 2016 (Table 2).

Table 2: Physical and inorganic parameters concentration of the SGRNK groundwater samples																
Well No.	рН	Conds- ctivity µS/cm	TDS mg1	Na mgʻi	K mg/	Ca mg1	Mg mgʻl	NH ₂ mgʻl	ପ mg1	Br mgʻl	SO4 mgʻl	F mg1	NO3 mg1	PO4 mg1	Fe mg1	Ma mg1
NP01	7	1090	742	196	4.4	62	0.45	0.28	275	0.22	55	0.17	2.4	0.09	-	-
P08	75	1290	920	172	9	88	58	0.5	232	0.15	204	0.68	0.3	0.05	-	-
NP03 P07R	7.4 7.6	4941 3370	3755 2516	1025 579	9	49 174	2.1 339	1.52 0.69	1220	22	659	0.56	3 0.3	0.53	ND ND	ND ND
P06R	7.1	4958	3720	688	5	241	622	0.45	1120	2.32	688	0.54	2.4	0.09	ND	ND
F05R	75	4931	3672	693	16	420	82.6	0.3	1126	3.66	922	0.38	1.7	0.38	ND	ND
F04R	79	1315	917	130	14	136	13.7	0.7	210	0.04	170	0.17	2.1	0.02	ND	ND
P03R	7.8	1450	920	244	7	13	6	1.25	255	0.06	194	0.28	0.6	0.02	ND	ND
F02R	75	581	415	99	5	6	9.55	0.63	124	0.02	15.2	0.1	0.7	ND	ND	ND
P01R	7.6	840 1220	572	136	9.2	11.2	4.5	0.56	175	0.06	18 122	0.32	1.2	0.23	ND	ND ND
P09R	7.7		892	243	9	13	11.6	0.68	290	1.22		0.41	0.5	0.21	ND	
P68RA P68RB	75 7.6	476 842	327 902	45 121	5.6 18	36.6 59	5.78	1.28	119 142	0.07	4.2	0.5	2.5	0.09	ND ND	ND ND
P68RC	77	423	276	53	10	31.2	12.6	0.79	89	0.06	56	0.03	1.5	0.09	ND	ND
FSOUB	75	854	560	92	8.8	34.5	6.7	0.5	102	0.13	72	0.7	3	0.41	ND	ND
FS OU	7.4	753	596	89	14	66	12.8	0.45	122	0.52	96	0.68	3.2	0.32	ND	ND
FS1UB	75	903	618	102	23	34	5.9	0.68	132	1.02	66.9	0.95	3.8	0.07	ND	ND
F81U	7	2415	1642	277	7.1	108	43.8	0.44	451	1.32	109	0.88	4.1	0.09	ND	ND
P24U	6.8 7.6	8253 1650	6120 1045	668	22	122.0	65.6	0.89	2836 176	6.32	90.7	0.59	1.8	0.09	ND	ND ND
P62U P18	7.0	5644	3520	126 890	15	129 120	1.8 19	0.66	1320	3.21	220	0.65	0.5	0.06	ND	ND
P18 P27U-1	6.8	9364	5216	1042	22	620	156	0.59	1320	1.32	962	0.08	0.7	0.12	ND	7.5
P27U	69	11998	6154	1245	21	857	130	0.38	2375	3.62	1422	0.92	1.6	0.09	0.11	7.5
P2 8UA	7.1	1272	890	143	12	69	342	0.2	195	23	163	0.5	0.5	0.08	ND	ND
P2 8UB	10	2955	1672	352	68	210	4.3	1.7	851	2.11	244	0.48	0.1	0.08	ND	ND
P2 8UC	7.7	3295	1762	535	15	63.8	30	1.08	612	1.4	242	0.41	0.1	0.05	ND	ND
PS3U	7.6	4028	2230	485	14	220	31	0.55	842	2.32	302	0.3	0.1	0.08	ND	ND
PS 3UA PS 3UB	7.4	3072 2685	2340	702 510	8	130 63	245 14	0.32	962 722	3.21 2.65	291 142	0.75	2.1	0.01	ND D	5.8 ND
F32U	72	1953	1473	253	12	143	243	0.75	284	1.52	247	0.38	1.5	0.03	ND	ND
FS 2UB	7	2408	1760	220	ii	204	41.8	0.81	514	1.32	95	0.44	0.7	0.05	ND	ND
P19	7.1	1591	1105	188	13	126	23	0.92	252	1.22	236	0.59	1.1	0.13	ND	ND
P58UB	72	2196	1560	183	12	169	38.8	0.49	575	0.62	74	0.32	2.1	0.09	ND	ND
P58UA	72	1765	1340	275	6.1	81	4.7	1.44	422	1.22	144	0.61	1.5	0.38	ND	ND
P59UB	73	1605	1094	322	12	47	495	0.48	342	0.09	92	0.92	0.9	0.05	ND	ND
P59UA	7 69	1532 1563	1120 1075	287 163	4.2 8.9	36.9 75	18.6 18.6	0.2	409 282	0.01	88 141	0.68	1.3 1.2	0.1	0.05 ND	ND ND
P60UA P60UP	7.1	2193	1549	476	7.6	29	1.41	1.49	632	0.12	56	0.30	2	0.01	ND	ND
P61UP	73	2051	1503	356	6.5	120	2.7	1.08	423	2.11	223	0.71	0.9	0.12	ND	ND
P61UA	69	1852	1304	287	4.2	48	0.88	123	412	1.32	90	0.81	1.5	0.01	ND	ND

In addition, the results of the seasonal changes of groundwater quality and some trace elements concentration pre and post the rainfall season are given in Table 3.

It is worth to mention that, Cluster analysis was used in this study, as it comprises of a series of multivariate methods which are used to find true groups of data related to basic hydrogeochemical analysis results (Table 2). In clustering, the objects are grouped such that similar objects fall into the same class (GAD 2001 and Danielsson et al., 2013). One of the benefits of the hierarchical method of cluster analysis, which is used in this study, is the advantage of not demanding any of prior knowledge of the number of clusters, which the nonhierarchical method does. A review by Sharma 1996 suggests Ward's clustering procedure to be the best, because it yields a larger proportion of correct classified observations than do most other methods (Davis, 1986, Gardner et al., 1990: Prethvivaj and Prakash, 1990). As a distance measure, Euclidean distances method is used in this study. The cluster analysis was carried out with single linkage and Euclidean method, firstly on the non-transformed input data matrix of all records. The results are given as R-mode dendrograms.

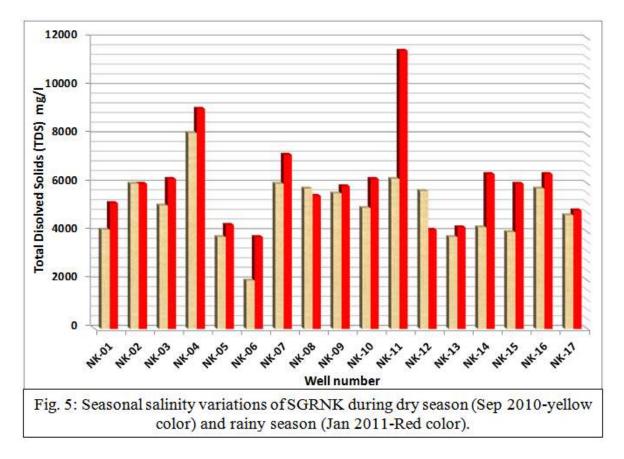
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post ra																			
SN WellN 1 P30UE		a Na 92	Na" 69	Diff -23	CI 102	CI* 169	Diff 67	F 0.7	F" 0.54	Diff -0.16	CaCO3 125	CaCO3 66	Diff -59	SO4	SO4'	Diff	PO4	PO4"	Diff
2 P30U	1	89	135	46	122	192	70	0.68	0.41	-0.27	105	110	5	1					
3 P31UE 4 P31U	4	102	63	-39	132	96 450	-36	0.95	0.78		175	135	-40	-					
4 P31U 5 P24U	1	277 668	184 1418	- <u>93</u> 750	451 2836		-1 277	0.59	0.43	-0.16	222 159	110 260	101	1					
6 P62U	1	126	229	103	176	374	198	0.63	0.91	0.28	210	60	-150	1					
7 P61UA 8 P61UE		287	213	-74	412	367	-45 -250	0.81	0.73		245	190	-55 44	-					
8 P61UE 9 P60UE		356 476	69 169	-287 -307	423 632	173 285	-230	0.79	0.05		116 240	160 180	-60	1					
10 P60UA	_	163	309	146	282	425	143	0.56	0.28	-0.28	180	290	110	I					
11 P59UA 12 P59UE		287	301	-217	409 342	409 272	-70	0.68	0.51	-0.17	272	240	-142	{					
13 P58U/		275	310	-217	422	448	26	0.61	0.21	-0.4	120	80	-40	1					
14 P58UE	3	183	142	-41	575	238	-337	0.32	0.38		152	200	48						
15 P19U 16 P32UE		188 220	195 175	-45	252 514	177 461	-75 -53	0.59	0.12	-0.47 -0.23	110 213	410	-103	1					
17 P32U		253	337	84	284	1099	815	0.38	0.31	-0.07	210	160	-50	1					
18 P33UE 19 P33UA		510	1842	1332	722	3600	2878 2138		0.71		170	290 270	120 160	ł					
20 P33U	1	702 485	1836 496	1134	962 842	3100 1120	2138	0.75	0.93	0.18	110 202	2/0	-102	1					
21 P28U0		535	553	18	856	920	64	0.41	0.35	-0.06	198	70	-128						
22 P28UE 23 P28U/		352	419	-23	851 195	755 177	-96 -18	0.48	0.05	-0.43	166	310	-91 176	{					
24 P27U	1	1245	1231	-14	2375		-415		1.17		273	410	137	1					
25 P27U	4 _	1042	1007	-35	1864	1820	-44	0.08	0.05		349	380	31	-					
26 P18U 27 P26U	-†≴	<u>890</u> 146	893 111	-35	1320 91	1725 189	405 98	0.43	0.27		170 109	210 150	40	ł					
28 P29U	R	360	404	44	547	761	214	0.5	0.18	-0.32	134	120	-14	1					
29 P63U 30 P49UA		<u>935</u> 487	1690 563	755	1702 975	2463 1276	761 301	0.72			167 213	60 190	-107	{					
31 P49U	12	715	665	-50	1276	1271	-5	0.09			256	30	-226	1					
32 P49UE		852	1553	701	2233		245	0.8	0.3		135	160	-153	-					
33 P25UE 34 P25U/		<u>625</u> 532	2015 427	1390 -105	1152 780	4325 836	3173 56	0.3	0.28		273 295	120	-165	1					
35 P20U	1	328	309	-19	332	488	156	0.47	0.34	-0.13	130	110	-20	1					
36 NP10 37 P23U	+	262	259	-3 379	320 1772	322 2410	2 638	0.71 0.68	0.6		124 156	<u>90</u> 60	-34 -96	{					
38 P50UE		13400		-4168	1320		405	0.64	0.79	0.15	210	70	-140	1					
39 P50UA 40 P50UA		12200		4302	<u>01</u>	189	98 214	0.05	0.05		236	70 60	-166	-					
41 P51UA		27700	9636	-18064 133	547 957	761	142	0.29			190 175	190	15	1					
42 P51UE		846	1108	262	1065	1612	547	0.01	0.02	0.01	175	100	-75						
43 P52UA 44 P52UE	-	<u>471</u> 288	361 498	-110 210	407 494	473 1117	66 623	0.36		-0.15 0.01	125 217	110 90	-15	ł					
45 NP12	1	430	258	-172	525	342	-183	0.92	0.48	-0.44	258	120	-138	1					
46 NP13	-	346	263	-83	563	325	-238	0.3	0.22		181	140	-41	-					
47 NP14 48 NP1	4	370 652	245	-125	425 835	345 384	-80 -451	0.48	0.19	-0.29	166 150	120 90	-46	4					
49 NP1:	;	110	56	-54	98	76	-22	0.11	0.05	-0.06	212	140	-72						
50 NP10 51 NP17	1	51 114	53 61	-53	63 173	29 90	-34 -83	0.05	0.05	0	109 109	120 160	11 51						
52 P53U 53 P53UA		233 102	131 147	-102 45	159 124	120 355	-39 231	0.26	0.17	-0.09	210 170	130 42	-80 -128						
54 P53U	Δ.	485	765	280	132.0	1488	168	0.33	0.17	-0.16	228	130	-98						
55 NP18 56 P17U		299	157	-142	332 159	242	-90 -37	0.09	0.22	0.13	140 84	130 160	-10 76						
57 P17U	A	75	124	49	119	210	91	0.22	0.09	-0.13	122	208	86	600	005	4.50	0.11	0.04	0.12
58 P47R 59 P47R		582	<u>591</u> 609	9 -111	376 177	821 978	445 801	0.05	0.09	0.04 -0.16	202 135	120 120	-82 -15	683 1201	225	-458 -1098		0.24	0.13
60 P48R	1	749	683	-66	421	355	-66	0.07	0.02	-0.05	110	210	100	929	890	- 39	0.25	0.39	0.14
61 P48R1 62 NP02	—	655	600	-55	301 375	837 386	536	0.21	0.1	-0.11	167	200	33 41	1007 145	311 36	-696		0.42	0.05
62 NP02 63 P11R	. v	325	283	-42 -422	835	386 358	11 -477	0.13	0.05	-0.08 -0.47	149 150	190 190	41 40	340	89	-109 -251	0.11	ND	0.02 ND
64 P10		230	209	-21	185	322	137 -431	0.09	0.05	-0.04	212.3	160	-52.3	179 315	40 1705	-139 1390	ND 0.21	ND 0.11	ND -0.1
65 P46R	<u>1</u> 2	174	156	-439 -21	886 230	455 236	6	0.2	0.09	-0.11 -0.12	75 110	20 200	-55 90	97	226	129		ND	ND
67 P46R		144 117	129	-15	176	193	17 29	0.11	0.28	0.17	166	140	-26 -45	232 208	348 344	116 136	0.07	ND ND	ND ND
68 P46R0 69 P14	4	182	220	5 38	138 157	167 325	168	0.78	0.26	-0.52	185 210	140 120	-90	141	12	-129	0.12	0.17	0.05
70 P55R		371 730	351 538	-20	210	160 825	-50 303	0.49	0.2	-0.29	193 445	170	-23 -345	297 382	572 326	275	0.06	0.02	-0.04
72. P47R		569	581	12	823	830	7	0.12	90.0	-0.03	146	110	-36	623	354	-269	0.14	0.09	-0.05
Na: So conc. b																			

III. RESULTS AND DISCUSSION

3.3 Seasonality change in salinity

The seasonality change in groundwater salinity is illustrated in (Fig. 5). It shows a comparison between salinity measurements at the end of the dry season (during September 2010) and in the middle of the rainy season (January 2011). The results show that the dry season salinities are, on the average, $\approx 25\%$ higher than the rainy season with a standard deviation of $\approx 5\%$. This quality change is important to be taken into consideration if this groundwater reserve is to be utilized for emergency purposes.

Moreover, Physico-chemical parameters were analyzed for groundwater samples have been summarized in Table 2&3. These parameters include pH, EC in situ, TDS, major cations and major anions (Table2). In general, pH is a measurement of activity of the free, un-complexed hydrogen ion which may lead to precipitation, co-precipitation and sorption processes that alter the chemical composition and reaction rates.



PH ranges 6.4–10 (mean 7.73) in pre and 6.1–9 (mean 7.3) for post-monsoon. The pH resultant contour map (Fig.5 right map) showed that pH of the groundwater in Sabriva wells decreased by range 0.8-1.3 mg/l and by 0.3-0.8 mg/l for the other three localities. This depicts slightly alkaline nature of groundwater due to the influx of HCO₃⁻ ions in the groundwater aquifer from percolation of rainwater through soil (Alam et al. 2011& Priyanka et al. 2016). All samples fall within the recommended limit (6.5 to 8.5) for human consumption (WHO, 2011) except sample No.P28UB was unsuitable for drinking purposes. Total dissolved solids (TDS) values range from 276-6154 mg/l (classified as 10% only fresh water, 38% classed as slightly saline groundwater and about 52 % moderately saline groundwater according to (Konikow and Reilly, 1999; Rhoades et al., 1992). The TDS resultant contour map (Fig.6 right map) showed that no change in TDS of the groundwater in Sabriya wells after rainfall season but decreased in the other three localities by 200mg/l. Total hardness (mg/l as CaCO₃) of groundwater is ranging from 75 to 445 (mean 179) for pre- and 20 to 410 (mean 150) for post-monsoon. TDS (mg/l) ranges 512.8–11570 (mean 4618.6) in pre- and 992–11790 (mean 4906.9) in post monsoon. The low TDS content in the groundwater of the study area could be either a result of short residence time with the underground rocks or the slow weathering of Dammam Formation limestone terrain (GAD 1995). All the hydrogeochemical analyses data recorded in tables 2 and 3 were used in mapping the studied physical and inorganic parameters of the groundwater in SGRNK by applying Golden software Surfer 13 (Fig.6 to Fig. 11).

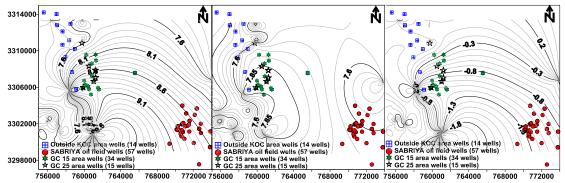


Fig. 6: pH contour maps before (left map), after rainfall (middle map) and the difference (right map).

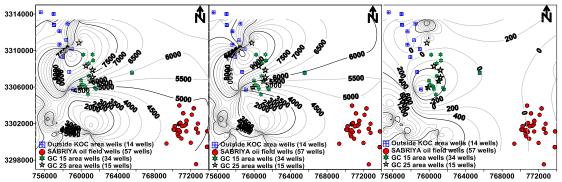


Fig. 7: TDS contour maps before (left map), after rainfall (middle map) and the difference (right map).

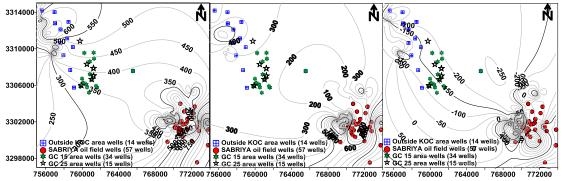
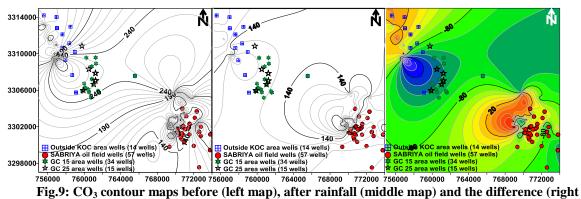


Fig 8: Na contour maps before (left map), after rainfall (middle map) and the difference (right map).



map)

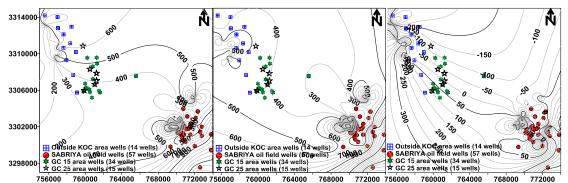
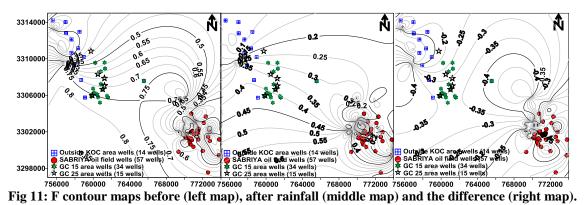


Fig 10: Cl contour maps before (left map), after rainfall (middle map) and the difference (right map).



Among major analyzed cationic concentrations (mg/l): sodium is the dominating ion ranges 51-27700 (mean 1157.2) pre monsoon and ranges 36-16502 (mean 963.7) post monsoon and followed by calcium ranges 6-1220 (mean 159) and 18-234 (mean 93.7), magnesium ranges 0.45-156 (mean 27.74) and potassium ranges 4.2-68 (mean 12.22) and 0.3-200.7 (mean 10.1) in pre- and post-monsoon, respectively. Among cationic concentrations (meq/l), Na predominates by constituting 44 and 40 % of total cations followed by Ca constitute 32 and 25 %, Mg constitute 23 and 33 % and K constitute 2 and 1 % of total cations in pre- and post-monsoon, respectively. Moreover, Na resultant contour map (Fig.8 right map) showed that Na concentration decreased after monsoon by 50 mg/l in Sabriya groundwater samples while in the other three localities by 100 mg/l. In this groundwater system, Na is the dominant cation that exceeds the threshold of dominance (meq/l 50 %) in 31 and 23 % of the samples followed by Ca (5 and 2 %) and Mg (0 and 6 %) in pre- and post-monsoon, respectively. So, hydrogeochemistry reveals that the order of cation abundance is Na > Ca > Mg > K in both dry and wet seasons.

Among major analyzed anion concentrations (mg/l): carbonate is the dominant ion ranges 75 - 445 (mean 179.1) and 20 - 410 (mean 150). The CO₃ resultant contour map (Fig.8 right map) showed that the concentration of CO₃ in the groundwater of Sabriya depression increased by 20 mg/l after rainfall season while the other three localities samples decreased by 80 mg/l. Moreover, the chloride ion ranges 63 - 2836 (mean 626.7) and 29 - 4325 (mean 820.8) in pre- and post-monsoon, respectively. It is noticed from the resultant contour map of Cl (Fig.10 right map) that the concentration of Cl in groundwater of Sabriya wells increased after rainfall season by 50 mg/l but decreased in the other three localities by range 50-100 mg/l. Otherwise, sulphate ion (SO₄) ranges 179 - 1007 (mean 67.13) and 40 - 311 (mean 20.73) followed by phosphate ion (PO₄) ranges 0 - 0.37 (mean 0.02) and 0 - 0.42 (mean 0.028) while fluoride ranges 0.01-0.95 (mean 0.44) and 0.02-1.24 (mean 0.3) in before and after rainfall season samples respectively as shown in Table 3. Also, nitrate ion (NO₃) ranges 0.1 - 4.1 (mean 1.56).

Among the anionic concentrations (meq/l), Cl is the dominant anion that exceeds the threshold of dominance (i.e. meq/l 50 %) followed by HCO_3 in pre- and post-monsoon samples, respectively. Hydrogeochemistry reveals that the order of anion abundance is $Cl > HCO_3 > NO_3 > SO_4 > CO_3 > F > PO_4$ in pre- and post-monsoon. In general, mineral weathering, dissolution and base-exchange processes control the levels of ionic concentrations in groundwater. The spatial distribution of total hardness in the investigated area clearly shows that high levels of hardness confined mainly to extreme southwest and few patches are found in central and northeast part in both pre and post monsoon (Table 2&3). High Ca, Mg and Cl in groundwater are the probable reason for the hardness in basin. Nitrate spatial distribution shows wide variation in its concentration infers point and non-point sources (Priyanka Patel et al. 2016). High nitrate contamination is

predominant in southwest part which consists of fast urbanizing area. Poor sewerage, leakage of human excreta from septic tanks and locally unmanaged solid waste disposal sites could have result in slug like motion of water during infiltration in groundwater (GAD et al. 2015). Few patches of nitrate pollution have also been observed in the central and northeast part of Umm Al-Aish might be due to high agricultural activities and domestic sewage. Application of N-fertilizers on irrigation land as crop nutrients along Abdali area may be responsible for nitrate pollution in the groundwater due to leaching by applied irrigation water.

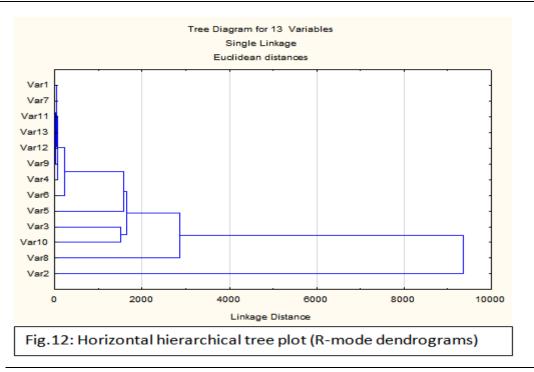
Moreover, two main groups of ion relationships can be concluded. Group1 which have ion dominance Na > Ca > Mg and $Cl > SO_4 > HCO_3$, where rNa/rCl is less than unity. This group (92% from 40 samples) characterizes groundwater samples which were collected from Al-Raudhatain and Umm Al-Aish areas (Table2). Group2 which have ion dominance Ca > Na > Mg and $Cl > SO_4 > HCO_3$ for samples No. PO4R and P24U extracted from Outside KOC area while ion dominance of Ca > Na > Mg and $SO_4 > Cl > HCO_3$ is characterizing to sample No. P62U related to Sabriya area.

According to salinity level, the water type is mainly freshwater type. On the other hand, the hydrochemical features of this group in the SGRNK presents Outside KOC area and Sabriya area which indicates marine salts contamination from the marine deposits of the neighboring catchment area and aquifer matrices. There is also a possible contamination for the SGRNK in Sabriya area with marine salts from the underline fractured limestone aquifer. This is confirmed by the assemblages of hypothetical salts combinations. Otherwise, the water composition of the SGRNK consists of NaCl, CaSO₄, Na₂SO₄ and CaCl₂. Whereas CaSO₄ and NaCl are the predominant water types in the brackish groundwater fields (Al Ruwaih and Ben-Essa 2004), in the freshwater fields, the principal water types are Ca(HCO₃)² and NaHCO₃. According to Kwarteng et al., 2000, the difference in water types is due to cation exchange process whereby the water quality changes from NaCl to CaCl₂ water type. The interaction of freshwater with a marine aquifer results in the loss of Ca⁺² and the formation of NaHCO₃ and Na₂SO₄ water types (GAD 1999).

Consequently, the origin of solutes was studied based on the Ca/Mg ratio. It is used to determine the sources of Ca and Mg into the groundwater. Ratio of 1, indicates dolomite dissolution, 1–2 indicate calcite dissolution dominance and >2 reflects an effect of silicate minerals (Raju et al. 2015). The Ca/Mg ratio varies 0.23-5.31(mean 1.63) and 0.16-2.73 (mean 0.86) (Table 2) in before and after rainfall season, respectively. 44 and 30 % groundwater samples depicts calcite weathering dominance, and 25 and 67 % shows dolomite weathering dominance in the study area in pre and post-monsoon season, respectively (Priyanka Patel et al. 2016). 31 and 3 % of the samples showing >2 Ca/Mg ratio which indicates silicate weathering is dominant process for the contribution of Ca and Mg ions in pre- and post-monsoon, respectively.

As a general, the chemical inorganic pollutants in the groundwater is discussed through measurement of trace elements concentration including Fe with minimum value (0.01 in wells No. P51UB and P52UB in Sabriya area) and maximum value of 0.95 in well No. P31UB as shows in (Table 3).Compared to international standards (WHO, 2011) to assess high concentrations of trace elements that could affect human health in the hypothetical case that impact shallow groundwater used for drinking water purposes. It is worth to mention that the most cations and anions are decreased in concentration after rainwater season which reflects a good chance for agriculture and sustainable urbanization development.

In addition, the identification and interpretation of Cluster analysis results (Fig. 12) were concluded. The rotated factor was also computed. Based on these steps, four statistical factors may be extracted as following (Fig. 12):



<u>Factor 1:</u> is the main factor and characterized by highly positive loading with Na, and SO_4 . This may attributed to the water rock interaction and municipal wastewater discharge and presence of surface saline soils, evaporation, agricultural activity and wastewater.

<u>Factor 2:</u> is highly positive loading with pH, NH₃, F, NO₃ and PO₄ due to agricultural activity, fluoride and silica enrichment and wastewater.

Factor 3: is highly positive loading with K and Pr due to CO₃-pH relation and industrial wastes.

Factor 4: represented by TDS, Ca, Mg and Cl which are independent variables.

IV. CONCLUSION AND RECOMMENDATIONS

The groundwater of the study area are slightly alkaline in nature. Majority of the groundwater samples are moderately hard to very hard waters in both before and after the rainfall season. In this groundwater system, Na (31 % of samples in pre- and 23 % in post-) and Cl (39 % of samples in pre- and 55 % in post-) are dominant ion that exceeds the threshold of dominance (i.e. meq/l 50 %). Ca/Mg molar ratio signifies calcite weathering as dominant source in pre- and dolomite weathering in post-monsoon for the ionic constituent in the groundwater. The spatial distribution of hydrochemical facies shows that majority of the area is dominated by Na–Cl facies. Deep saline water of Dammam aquifer upconing and wastewater infiltration is the primary factors along with agricultural return flow and sea water intrusion which play momentous role in increasing salinity of the groundwater. According to Cl classification, brackish water is distributed all over the study area followed by the fresh brackish water. Four clusters were inferred from the principal component analysis noticeably infers the water rock interaction and municipal wastewater discharge in PC1, agricultural fertilizer input in PC2, CO3-pH relation in PC3, Salinity and dolomite weathering beside silicate weathering is dominant process in PC4 as major factors in the groundwater system. These findings reveal that groundwater is less polluted with various natural and anthropogenic activities and majority of the samples are good to permissible for drinking purpose. The groundwater usage for domestic purposes has been limited due to the excessive permissible limit of TDS, TH, Na, HCO₃ and NO₃ in the study area.

The study area is suitable for agricultural and urbanization development in case of good planning especially with respect to the fresh groundwater lenses in SGRNK. Also, long-term monitoring of both groundwater levels and quality in the future will be required to assess the threat to groundwater reserves and to adopt appropriate mitigating actions. Also, the flow of oil from the oil wells in the vicinity, damaged during the 1991 Gulf War, and subsequent use of seawater for extinguishing the oil fire have contaminated the soil and infiltrating rainwater has carried these pollutants to the relatively shallow groundwater in these areas. It must be recommended that these environmental problems related to the damaged during the 1991 Gulf War are badly in need of treatment through continuous projects as fast as possible.

ACKNOWLEDGMENTS: The author acknowledges Ministry of Electricity and Water (MEW) for financial support and for supplying the data from the Project entitled: GROUNDWATER QUALITY MONITORING PROGRAM PHASE1 (2016-17) and PHASE2 (June 2017-July 2017). The author is thankful to NAPESCO team work for their constant help and logistic support for groundwater samples collection and analysis during field work of the project.

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Hamoud N. Alalati"The Effect of Seasonal Rainfall on Groundwater Quality in North Kuwait. "International Journal Of Modern Engineering Research (IJMER), vol. 08, no. 07, 2018, pp.69-82